

## Estimation of Carbon Stock in the Regenerating Tree Species of the Intact and Disturbed Forest Sites in Tanzania

Elly Josephat Ligate<sup>1,2\*</sup>, Can Chen<sup>3</sup> and Chengzhen Wu<sup>1,3\*</sup>

<sup>1</sup>College of Life Sciences, Fujian Agriculture and Forestry University, Fujian 350002, P. R. China.

<sup>2</sup>Department of Biosciences, Sokoine University of Agriculture, P.O.Box 3038, Morogoro, Tanzania.

<sup>3</sup>College of Forestry, Fujian Agriculture and Forestry University, Fujian, 350002, P. R. China.

### Authors' contributions

This work was carried out in collaboration between all authors. Author EJJ designed the study, performed the statistical analysis and wrote the first draft of the manuscript. Authors CC and CW managed the analyses of the study. Author CW managed the literature searches. All authors read and approved the final manuscript.

### Article Information

DOI: 10.9734/IJECC/2018/42020

Original Research Article

Received 10<sup>th</sup> May 2018

Accepted 19<sup>th</sup> May 2018

Published 7<sup>th</sup> June 2018

### ABSTRACT

**Aim:** Estimation of carbon in the forests located in the coast of tropics is needed to support conservation and forest monitoring strategies. This study aimed at quantifying carbon stocks in the regenerating tree species of intact forest (IFS), disturbed by agriculture (ADS) and by livestock grazing sites (LDS) to understand the importance of coastal trees in carbon stocking as part of mitigating climate change impacts.

**Methodology:** Thirty-three independent measurements of tree carbon stocks were carried out on 33 tree families found in the coastal zone of Tanzania. The vegetation was inventoried by means of a floristic survey of the woody component across intact, crop agriculture and livestock disturbed land use sites. The biomass was then estimated by employing the existing allometric equations for tropical forests. Thereafter, the above ground stored carbon was quantified on the sampled tree species found in each land uses.

**Results:** The results showed that there were significant variations ( $p \leq .05$ ) of carbon stock values across species and land uses. The average carbon (Kg/ha) stored in the regenerated adult trees was 1200 in IFS, 600 in ADS, 400 in LDS. Saplings had 0.43 in LDS, 0.07 in ADS and 0.01 in IFS. Indeed, seedlings had the average of 0.41 in IFS, 0.22 in ADS and 0.05 in LDS.

**Conclusion:** These findings show that crop-agriculture highly affects the regeneration potential of trees, biomass accumulation and carbon stock than livestock grazing. To restore carbon storage

\*Corresponding author: E-mail: ligateelly@yahoo.com, fafuwcz@163.com;

potential of coastal tropical forests, crop-agriculture must be discouraged, while livestock grazing can be integrated in forest management. Indeed, further studies are required to gauge the integration levels of any anthropogenic activities, so that the natural capacity of coastal tropical forests to regenerate and stock carbon is not comprised further.

*Keywords: Carbon sink; carbon source; sequestration; land uses; regeneration.*

## 1. INTRODUCTION

Life on earth depends largely on forest ecosystems and their services [1,2,3]. In nature, forest ecosystems are the major terrestrial reservoirs of carbon in the form of plant biomass and soil organic matter [4,5,6,7]. These ecosystems are among the locally and globally recognized sources or sinks of carbon in the remaining or regenerating forests [2,5,8,9]. Forests play crucial roles in regulating the global biogeochemical cycles [7,8,10,11,12]. Indeed, forest ecosystems play important roles in reducing the atmospheric carbon dioxide and hence regulating climate change [13,14,15]. Thus, forests are among the vital components of the global ecosystems in addressing climate change, the most pressing issue in the world today [8,16,17]. Although forests are important sources or sinks of carbon, they are frequently affected by human activities pressure [18,19,20]. Regionally and globally, human activities disturb forest ecosystems through land cover and land use (LCLU) changes, hence causing forest ecosystems to function as carbon sources rather than sinks [21,22,23,24,25]. These activities have contributed to the introduction and development of secondary forests in the tropics [26]. Human activities cause land cover and land use changes that pose challenges on the capacity of forest to regenerate, function, and offer various ecological services including the capacity to function as carbon sources and sinks [27,28,29]. However, there is little information about the amount of carbon in regenerating forests ecosystems along the tropical coastal forests (in this study referred as tropical coastal forests). Therefore, it is important to estimate carbon stocks of the regenerating species for understanding their contribution in the global carbon stock and in addressing climate change [30].

Carbon loss in forests ecosystems are the outcomes of anthropogenic activities [8,25]. Deforestation and degradation of terrestrial forest ecosystems are the main factors for the loss of carbon in the tropical forests [25,31,32]. Deforestation, degradation, and poor forest

management reduce the capacity of forests ecosystems to store carbon [33]. These activities bring the so called anthropogenic causes of global warming [6,32]. The loss of carbon is based on the fact that disturbances affect the structure of forests including the type, size, age, species stand and species diversity, the parameters, which are directly associated in the storage of carbon in forest ecosystems [10,12,34]. Also, disturbances in forest affect the belowground carbon stock that include soil, litter and roots [32,35,36]. Although, documentation shows that the below ground carbon sink of trees harbors larger quantities of carbon, this sink capacity is limited by many factors such as the magnitude of historic carbon loss, higher rate of decomposition because of change in climate, and different land uses and management [23,37].

The land use forms such as cultivation and livestock grazing expose soils to loss of the sequestered carbon in the terrestrial forest ecosystems [20,23]. These activities disturb the capacity of the below ground carbon storage system, which stores the largest terrestrial carbon pool (i.e. storing more than double the quantity of carbon in vegetation or in the atmosphere) [38]. Unquestionably, crop-agriculture and livestock grazing are among the major activities contributing to forest LCLU changes in the tropics [19,22,39]. These activities contribute substantially to alter carbon storage in forest ecosystems [40,41]. It is clear that crop-agriculture and livestock grazing fail to support forest ecosystem sustainability and to restore the degraded ecosystems [42,43].

In order to allow regeneration in the disturbed and degraded ecosystems, different ecosystems management options are implemented, in which exclusion of anthropogenic activities are implemented in many parts of the world [44]. Exclusion is sought to contribute in allowing forests to regenerate naturally and thus many of the existing forests species in the tropics are secondary [45]. Existing studies have quantified the amount of carbon in various ecosystems. For example, carbon storage in grasslands ecosystems [40,44], carbon storage in the

tropical forests [34,35], land use changes and carbon emissions in terrestrial ecosystems [18,21]. Other studies include land use changes, carbon sequestration and biodiversity [17,22,31,46] and carbon storage in plantations [6,33], while allometric models for species located in the coastal zone of Tanzania have been presented in [43]. Across all these studies, it is indicated that trees play a big role in carbon storage. These studies show that an increase in secondary natural forest increases carbon storage above and below the ground [18]. Nevertheless, literature show that land use conversions affect the capacity of forests to store carbon [22]. To revert the trend of forest loss and impacts on carbon storage, regional and global efforts are increasing to implement land use changes that aim at restoring biodiversity and the degraded forest ecosystems [8,45]. These efforts promote re-growing of forest species, thus automatically facilitating regeneration process and creation of carbon sink [4,47]. In fact, the knowledge about forest and carbon interplays in the tropics provides information for management of these vital ecosystems, which dominate the role of controlling the global carbon cycle based on both carbon flux and the volume stored in these forests [48].

This study sets a baseline for future comparisons of carbon stock after exclusion of human activities bearing in mind that carbon sequestration increases with forest restoration age [22]. The information generated in this work, provide basic information to operationalize value to land managers and policy makers as they facilitate monitoring of tropical forest carbon dynamics and further motivation to conserve tropical forests for reducing net CO<sub>2</sub> emissions [2,25,33,35]. In the present study, we examined the variation and established the relationships between regenerating tree carbon storage across intact forests sites (used as control) and forests disturbed land use sites after exclusion of crop and livestock production. We focused on estimating carbon in the above ground biomass of seedlings, saplings and adult trees because the above ground carbon is stored in tree biomass [7,35]. Specifically, the study focused on analyzing the difference in carbon sink across intact and disturbed sites because different LCLU cause variation in the amount of carbon held in terrestrial ecosystems [14,18,48]. The following hypothesis was tested. Carbon storage differs between regenerating species in closed forests sites from the sites disturbed by crop-agriculture and livestock grazing. This work was carried out

to find the answer to the following question: How carbon varies across regenerating species of forest sites subjected to different land uses and management?

## 2. MATERIALS AND METHODS

### 2.1 Description of the Study Area

This study was conducted in the forest located along the coastal zone of Tanzania. The study area was Uzigua Forest Reserve (UFR) located in Bagamoyo and Chalinze Districts in the Pwani Region of Tanzania (Fig. 1). This forest is located within 100km from the Indian Ocean. Specifically, the forest is found between 60° 00' and 60° 15' and 38° 00' to 38° 15'E [49]. The forest is characterized by being affected by human activities mainly crop-agriculture, tree harvesting for charcoal and timber production, livestock grazing pressure and encroachment for human settlements. It is because of historical characteristics of these anthropogenic activities. Therefore, UFR was purposely selected for this study.

### 2.2 Inventory of Tree Species

Ground forest inventories were carried out to measure and identify tree species sub-categories (i.e. seedlings, saplings and adults) from IFS, ADS and LDS [50]. Trees population density, diameter at breast (dbh), and height were measured for determination of basal area and bio-volume of each species. These determined variables were used for allometric equation by relating wood volume to stem diameter at breast height [51]. A random selection of sites and the establishment of sampling plots were carried out after the stratification of the land use sites. Forty-five (45) quadrats of 25 m × 25 m size were laid down for collection of adult trees data, while nested plots of 2 m × 2 m (within the established 25 m × 25 m plots) were laid down for collection of seedlings and saplings data [52,53]. Stems with a diameter of ≥ 20 cm at breast height (dbh) (approximately 1.34 m height above the ground) were counted as trees. All the tree species with < 20cm diameter were considered as regenerates in the following subdivisions: (i) seedlings included only trees with < 0.40 m height and (ii) saplings included all the trees from ≥ 0.40 m to < 1m heights as adapted from [53]. Seedlings, saplings and adult trees were identified and recorded in each of the same established sampling plots across the sites as adapted from [34]. Photos of trees species were taken in the

field to verify the accuracy of the field plant identification.

### 2.3 Analysis of Tree Species Data

In order to quantify carbon from the regenerating trees, we adopted a non-destructive method in collecting tree species parameters and then computed the biomass and carbon for each seedling, sapling and adult trees [4,5]. From tree species checklists (i) number of live trees per unit area (N/ha), (ii) basal area (BA) of live trees (m<sup>2</sup>/ha), and (iii) volume of live trees (m<sup>3</sup>/ha)

were calculated following a methodology laid down by [54]. Computation of BA was carried by by  $BA = ((dbh)^2 \times \pi) / 4$  (Eq.1); where dbh = diameter at breast height and  $\pi = 3.14$ ; the volume was calculated as  $v = ghf$  (Eqn. 2); where v = volume estimation (m<sup>3</sup>/ha), g = basal area of the tree/seedling/saplings (m<sup>2</sup>/ha), h = height of the tree (m) and f = form factor (0.5). We used the form factor of 0.5 as an average for natural forest factor, which ranges between 0.4 and 0.6 [30,55]. The computation of these factors was done by ensuring that each land use class is represented [56].

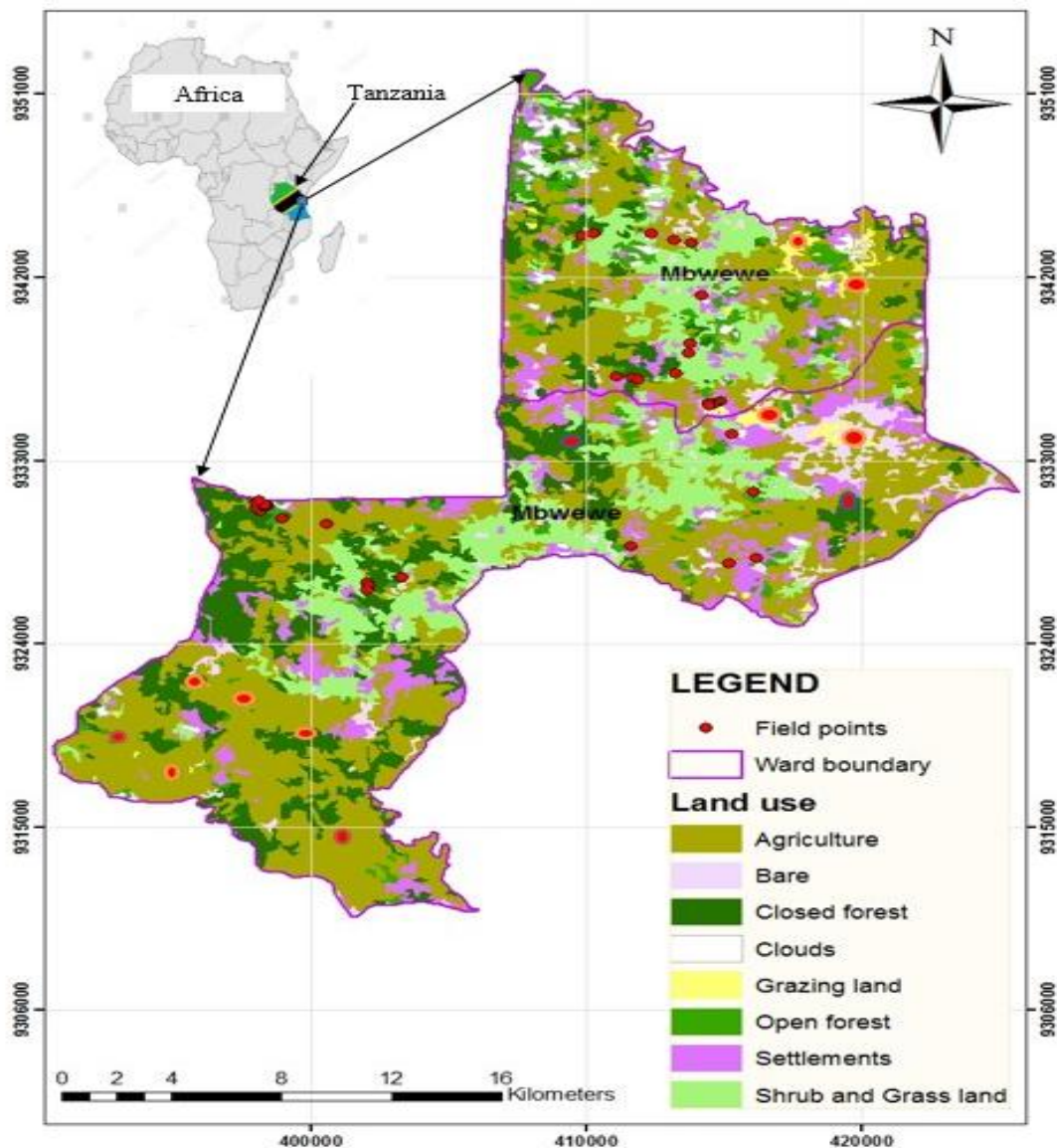


Fig. 1. A map of study area

## 2.4 Determination of Tree Biomass and Carbon Content

Tree biomass and carbon pools were determined using allometric equations from GlobAllomeTree platform (The international platform for tree allometric equations) [51,57]. We used the equations particularly developed for the tropical tree species as in [51,58]. These models were used in computing the above ground (ABG) and carbon stock per each tree species, on each sampling plots [4]. The AGB were estimated as  $AGB = V \times WD$  (Eqn.3); whereby V is the bio-volume and WD is the wood density for each tree species [33,57,59]. To maintain the non-destructive methodological approach (because we were not permitted to harvest any part or whole plant from the reserve, except photographing as showed in Table 1), the WD for each species were adopted from [43,60,61]. Carbon stock per each species in each sampling plots was estimated as:  $C = TB \times CF$  (Eqn.5), where by C is the carbon, AGB is the above ground biomass and CF is a carbon fraction of dry matter (default = 0.5), tonnes of carbon [62].

## 3. RESULTS

### 3.1 Tree Families and Species Studied

The number of species, which were recorded for carbon stock estimates were 33. These species

were from 14 families including Fabaceae (39%), Moracea (13%), Chrysobalancea, Combretaceae, Guttiferae), and Malvaceae (each by 6%), Asteraceae, Ebenaceae, Lamiaceae, Lauraceae, Meliaceae, Rhamnaceae, Rhizophoraceae and Rubiaceae each being represented by 3%. Throughout the figures, each tree species was represented in Arabic numbers as arranged in Table 1.

### 3.2 The Mean of Carbon Stock Across Tree Sub-categories

The mean of carbon stock (Kg/ha) across the tree sub-categories were  $1.22E3 \pm 101.59$ ,  $4.72E2 \pm 60.37$  and  $6.33E2 \pm 90.28$  for adult trees in IFS, ADS and LDS respectively. The mean carbon for seedlings was  $0.5 \pm 0.01$ ,  $0.22 \pm 0.03$  and  $0.41 \pm 0.05$  Kg/ha in the IFS, LDS and ADS respectively. The mean carbon in saplings was  $0.01 \pm 0.01$ ,  $0.07 \pm 0.01$  and  $0.43 \pm 0.04$  Kg/ha in IFS, LDS and ADS respectively.

There was a significant difference in carbon stock between adult tree in IFS and ADS with the mean variation of  $7.46E2 \pm 88.70$ , at  $t = 8.41$  and  $p < .001$ . There was a significant difference between carbon in adult tree found in IFS and LDS as indicated in the mean of  $5.84E2 \pm 157.65$ ,  $t = 3.71$ ,  $p < .001$  and the difference of carbon in adult trees in ADS and LDS showed a significant value of  $1.62E2 \pm 116.93$ ,  $t = 1.38$ ,  $p < .177$ .

Table 1. A list of tree species used in this study



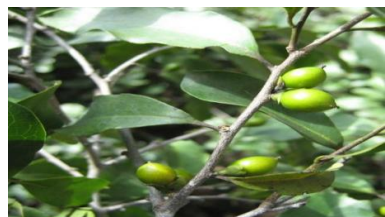
1. *Tamarindus indica*



2. *Afzelia quanzensis*



3. *Dialium holtzii*



4. *Diospyros abyssinica*



5. *Albizia versicolor*



6. *Tectona sp.*



**7. *Albizia gummifera***



**8. *Julbernardia globiflora***



**9. *Dalbergia melanoxylon***



**10. *Terminalia sambesiaca***



**11. *Milicia excelsa***



**12. *Allanblackia stuhlmannii***



**13. *Khaya anthotheca***



**14. *Terminalia superba***



**15. *Sterculia quinqueloba***



**16. *Artocarpus heterophyllus***



**17. *Baphia* sp.**



**18. *Xeroderris stuhlmannii***



**19. *Brachylaena huillensis***



**20. *Combretum schumannii***



**21. *Berchemia discolor***



22. *Brugueira gymnorhiza*



23. *Pericopsis angolensis*



24. *Combretum zeyheri*



25. *Brachystegia boehmii*



26. *Brachystegia spiciformis*



27. *Vangueria sp.*



28. *Pterocarpus sp.*



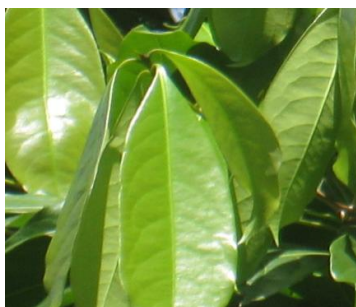
29. *Parinary sp*



30. *Ficus sp.*



31. *Newtonia buchananii*



32. *Ocotea sp.*



33. *Sterculia appendiculata*

In regards to seedlings, the difference of carbon showed higher values between seedlings in IFS and ADS with the mean variation of  $0.17 \pm 0.03$ ,  $t = 6.59$ ,  $p < .001$ , carbon in IFS and LDS variation was low with the mean value of  $0.36 \pm 0.04$ ,  $t = 8.02$ ,  $p < .001$ , while carbon in ADS and LDS had a mean difference of  $0.19 \pm$

$0.05$ ,  $t = 3.44$ ,  $p < .002$ . The mean difference for saplings between IFS and ADS was  $0.05 \pm 0.01$ ,  $t = 7.34$ ,  $p < .001$ . The variation between IFS and LDS saplings carbon stock showed the mean of  $0.42 \pm 0.04$ ,  $t = 10.75$ ,  $p < .001$  and that between ADS and LDS was  $0.36 \pm 0.04$ ,  $t = 8.86$ ,  $p < .001$ .

### 3.3 Saplings and Seedlings Carbon Stock across Species in IFS

Carbon stock varied across species. Higher carbon stock was recorded in saplings than in seedlings in IFS. Higher carbon stock was observed in saplings of *Azelia quanzensis*, *Brugueira gymnorhiza* and *Milicia excelsa* with carbon ranging between 0.05 Kg/ha to 0.19 Kg/ha, while the seedlings carbon stock was dominated by *Brachystegia boehmii*, *Diospyros abyssinica* and *Parinari* sp. at the range of 0.02 Kg/ha to 0.03 Kg/ha. Other species had low contribution to carbon across saplings and seedlings per ha as shown in Fig. 2.

### 3.4 Saplings and Seedlings Carbon Stock across Species in ADS

Unlike in the in IFS, ADS had low values of carbon stock across species and tree sub-categories. Saplings dominated the seedlings component as presented with species mainly *Berchemia discolor*, *Combretum schumannii*,

*Milicia excelsa* and *Sterculia quinqueloba*. Seedlings carbon stock was mainly contributed by *Brachylaena huillensis*, *Brugueira gymnorhiza*, *Dalbergia melanoxylon* and *Tamarindus indica*. Saplings and seedlings, which contribute largely to carbon stock had a stock ranging between 0.36 Kg/ha to 0.58 Kg/ha, while less values of carbon were recorded in other species as shown in Fig. 3.

### 3.5 Saplings and Seedlings Carbon Stock across Species in LDS

The trend of carbon stock in LDS differed across the species and tree categories. The stock of carbon was significantly contributed by saplings such as *Azelia quanzensis*, *Dialium holtzii*, and *Diospyros abyssinica* and *Tamarindus indica*. Seedlings carbon values were dominated by *Baphia* sp., *Brachylaena huillensis*, *Pericopsis angolensis* and *Tamarindus indica*. These species had values ranging between 0.50 Kg/ha and 0.93 Kg/ha. Other species had the mean carbon stock below 0.6 Kg/ha as shown in Fig. 4.

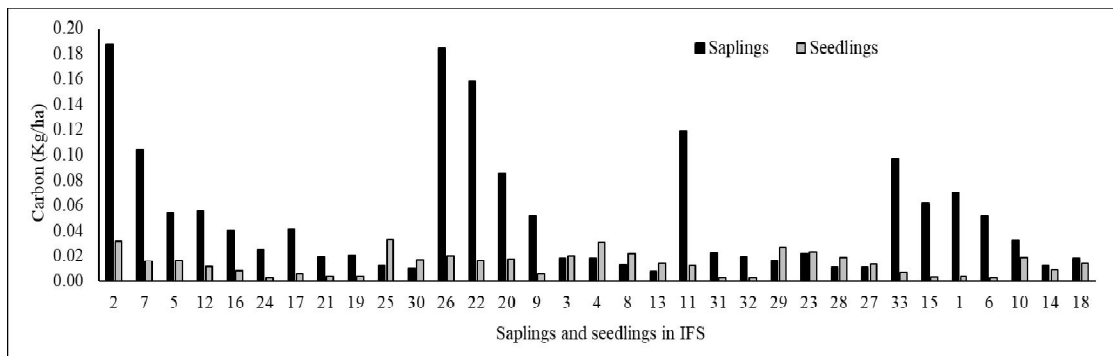


Fig. 2. Saplings and seedlings carbon stock in IFS

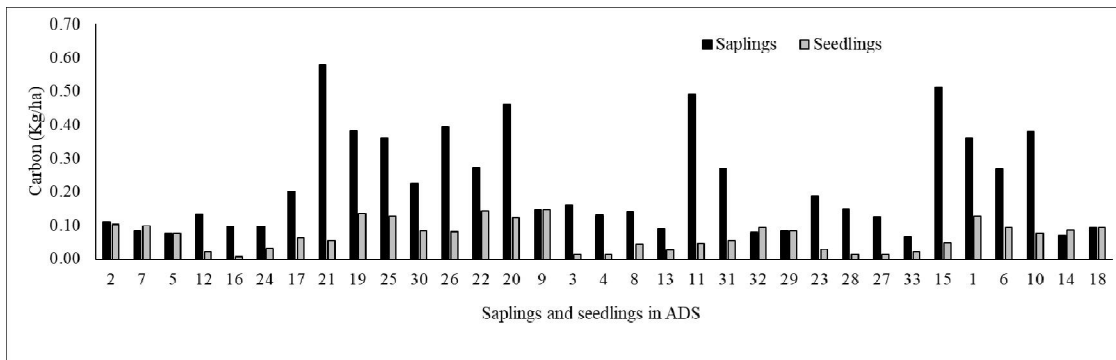


Fig. 3. Saplings and seedlings carbon stock in ADS



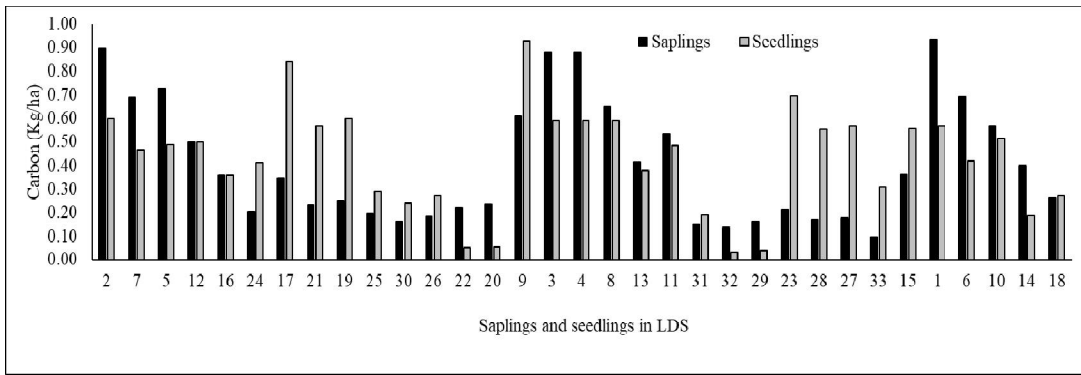


Fig. 4. Saplings and seedlings carbon stock in LDS

**3.6 Adult Trees Carbon Stock across Species in IFS**

Carbon stock in adult trees was higher in *Tamarindus indica*, *Allanblackia stuhlmannii*, *Baphia* sp. and *Parinari* sp. The carbon stock in these species ranged between 200 Kg/ha to 3000 Kg/ha. The lowest stock was in *Terminalia sambesiaca*, *Milicia excelsa*, *Berchemia discolor*, *Brugueira gymnorrhiza*, *Brachystegia boehmii* and *Vangueria* sp. with the stock value below 200 Kg/ha as shown in Fig. 5.

**3.7 Adult Trees Carbon Stock across Species in ADS**

In ADS, the highest carbon stock was recorded in *Tamarindus indica*, *Artocarpus heterophyllus*, *Baphia* sp. and *Parinari* sp. These species had carbon values between 200 Kg/ha and 1500 Kg/ha. The lowest values were recorded in *Brugueira gymnorrhiza* and *Vangueria* sp. with the carbon stock of less than 200 Kg/ha (Fig. 6).

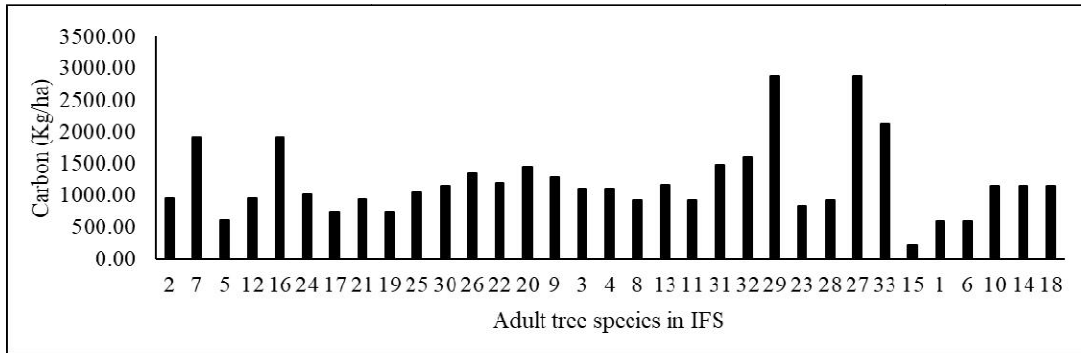


Fig. 5. Adult trees carbon stock in IFS

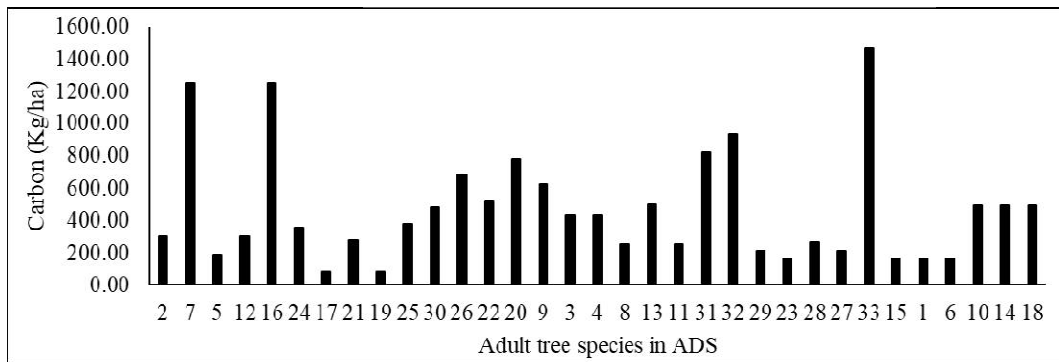


Fig. 6. Adult trees carbon stock in ADS

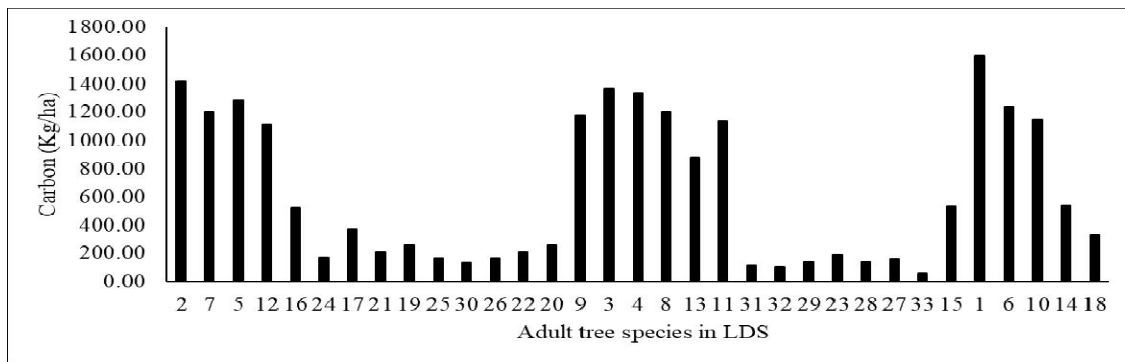


Fig. 7. Adult trees carbon stock in LDS

### 3.8 Adult trees Carbon Stock across Species in LDS

In LDS, the highest stock of carbon was in *Tamarindus indica*, *Dialium holtzii*, *Artocarpus heterophyllus*, *Baphia* sp. and *Parinari* sp. with carbon stock between 200 Kg/ha to 1500 Kg/ha. The lowest carbon stock was recorded in *Berchemia discolor*, *Brugueira gymnorhiza* and *Vangueria* sp. each with carbon below 200 Kg/ha (Fig. 7).

## 4. DISCUSSION

This study based on thirty-three (33) trees to represent some species, which had high appearance in the study area like the criterion used in [43]. Unlike in many previous studies, this presented work shows carbon stock in different tree sub-categories (seedlings, saplings and adult trees) across three land uses (intact, crop-agriculture and livestock disturbed sites). Therefore, in the discussion, there are some limitations in some data and information (about carbon stock in saplings and seedlings) for comparison of our findings. However, the findings in this work set a baseline for future comparison of carbon stocks in the regenerating trees (i.e., saplings and seedlings).

### 4.1 Carbon Storage across Land Uses

Findings in this study showed a variation of carbon stock across intact forest, crop-agriculture and livestock disturbed sites. This variation shows that human activities (crop-agriculture and livestock grazing) affect forest structure. In turn, these activities affect the potential of trees to function as carbon sink and store on particular land uses. From the findings of this work, it indicates that different management of different

land uses and disturbances affect carbon storage in the vegetation component of forests ecosystems [59]. Carbon stock in the regenerating species indicates that disturbances affect vegetation, but after some restoration measures such as exclusion of crop-agriculture and livestock grazing, there are some potential values to rejuvenate forests and restore the capacity of these ecosystems to store biomass carbon as supported by [44]. The potential to function as carbon stocks is mainly based on the capacity of the land use to permit regeneration and growth of trees. The variation of carbon stored in a particular land use shows that it is not only the density of species that determines carbon amount, but the capacity of species to store carbon, which differs from one species to another as determined by many factors such as heights, agreeing to the findings in [18]. The variation of carbon storage across trees sub-categories was expected in this study because trees categories differ in heights and diameters, and these factors hold important implications for carbon storage potential in tropical forests [63]. The computed carbon stock across the study sites indicates that coastal forests play important role in ecosystems services such as carbon storage [43], but different land management affect the tree growth parameters and carbon storage potential.

### 4.2 Carbon Stock between Intact and Disturbed Sites

Across the study sites there was less carbon in crop- agriculture regenerating species agreeing with [47]. The intact forests sites had higher carbon stock than crop- agriculture and livestock disturbed grazing sites. The higher amount of carbon in intact forest sites indicates that protection or allowing natural regeneration to

take place contributes to store above ground carbon stocks [18]. The higher carbon stock in intact forest sites is within the average range reported for adult trees in [9,64,65]. Low carbon stock in crop-agriculture and livestock grazed sites, shows that disturbances affect regeneration of trees, and hence there is low carbon storage in the tropics supporting [66] findings. Low carbon in disturbed sites is a result of low density of adult trees. Interestingly, disturbed sites had carbon potential in saplings and seedlings, which equally compares to the average quantity of carbon stocks in bushland and grasslands of the tropics [56,57].

#### **4.3 Carbon Stock within the Disturbed Sites**

The amount of carbon in the regenerating species of the livestock grazed sites differed slightly from that on crop-agriculture sites. Although both livestock grazing and crop-agriculture are associated with vegetation disturbances in forest ecosystems; these activities affect the above ground forest biomass and carbon stocks differently [23,66,67]. From the study sites, it is obvious that the impacts of livestock grazing are somehow less than those caused by crop-production because livestock grazing is selective and leaves some species unaffected unlike crop-agriculture. Indeed, the amount of carbon recorded within these two land uses shows that, these lands have the potential to regenerate forest trees, contributing to conservation of trees within the previously disturbed sites, in turn improving the storage of carbon [39,37]. Basically the carbon stored in the regenerating tree species is a sign that coastal forests have high capacity of resilience of carbon stocks that are can be enhanced through conservation and restoration [68].

#### **4.4 Carbon Stock across Tree Sub-categories**

The general trend showed a substantial increase in carbon stock across tree sub-categories and land uses. Carbon stock was less in seedlings but higher in saplings and adult trees across the land uses. The variation of carbon stock across tree sub-categories indicates that as trees grow accumulates higher carbon than the regenerating seedlings and saplings. This view supports the observations in [48]. Carbon variation across seedlings, saplings and adult trees shows that the regenerating seedlings and saplings play carbon storage function not like the role played

by mature and old-growth natural forests [69]. The estimated quantity of carbon in the seedlings and saplings confirms the potential of tropical forests to regenerate and store carbon after conservation measures [37]. Carbon storage in seedlings and saplings shows that the young forests constitute carbon storage of coastal forests like many other tropical forests [69]. The low variation of carbon in seedlings and saplings (regenerating trees) shows that carbon pools and regenerating species have different recovery rates [70,71]. The carbon in disturbed sites shows that disturbance lowers carbon content in the ecosystems, and it might take long time for the disturbed sites to rejuvenate and gain higher levels above ground biomass and carbon stock potential [71,72]. Specifically, this variation shows the contribution of the regenerating tropical forests located in the coastal zone in reducing CO<sub>2</sub> in the atmosphere. The low variation of carbon in seedlings and saplings across the land use sites suggests that degraded forests and abandoned agricultural lands have the potential to recover as well as play the role of carbon and biodiversity values, if they are left to regenerate naturally [72].

#### **4.5 Carbon Stock across Different Species**

The variation of carbon stock across different species in this study shows that different plant species have different capacity to sequester carbon during photosynthesis supporting the findings in [32]. Although in this study we have used the generalized allometric equations to quantify carbon stock for all the thirty-three species, interestingly, the computed values of carbon in our study area are within that reported in other studies like [9,25,73,65], but they are contrary to the values reported in [32]. These contradicting findings might result from variation of species, location, age of the tree and methods of quantifying carbon stocks [32]. It is possible that the variation, which is between our work and the existing literature, would have been counterbalanced if we had used the destructive methods of carbon assessment across the species. However, the variation established across the study sites and tree species suggests that farming and livestock grazing have impacts on forest carbon stocking [9,66]. In this study, it shows that exclusion of human activities in the tropical coastal forests facilitates natural regeneration, and thus improving carbon stocking. Therefore, the regenerating species play important role in carbon storage like many

other natural forests agreeing to the documentation in [74].

#### 4.6 Carbon Stock and Its Implications on Climate Change

The interplay between forest disturbances, regeneration, carbon sources or stocking and climate change is complex because climate change is both a cause and an effect of forest change [74,75]. The quantified carbon stock across land uses and tree sub-categories is important in understanding the role of regenerating forests in addressing climate change mitigation [76]. Our findings show lower carbon stock per unit area agreeing the findings in [57] but contrary to [77,78,79]. This controversy shows that forests disturbance in Tanzania is high and continue to be a challenge in addressing global efforts to mitigate climate change [57,79]. Lower carbon stock in the disturbed sites implies that disturbances affect the potential of forests to store carbon. However, these findings highlight that there is some carbon stocking potential in some of the regenerating trees for carbon sequestration and climate change mitigation. Therefore, converting the disturbed sites into forests may increase carbon sequestration as some tree species have good capacity to regenerate and play crucial role of carbon storage, a function, which is important in addressing climate change after disturbances.

#### 5. CONCLUSION

This study confirmed the hypothesis that carbon storage differs between regenerating species of the intact forest sites from crop-agriculture and livestock disturbed sites. The study concludes that there are significant variations of carbon stock values across the thirty-three species, tree-sub-categories and the average amount of carbon across the three land uses. These carbon stock variations are useful indicators that different land use management affect the potential of coastal forests to function as carbon sinks in addressing changing climate mitigations. Indeed, the higher quantities of carbon in adult trees of the intact forest sites than those found in the disturbed sites provide a useful information that disturbances that cause loss of forest trees results into forests to act as carbon sources rather than sinks. Therefore, it is important to promote restoration, protection and conservation of forest species to optimize carbon stocking benefits for sustainable management of coastal forest ecosystems.

#### COMPETING INTERESTS

Authors have declared that no competing interests exist.

#### REFERENCES

1. de Groot R, Ramakrishnan PS. Ecosystems and human well-being: Current state and trends, Chapter 17: Cultural and amenity services. *Ecosyst. Hum. Well-being Curr. State Trends*. 2005;457–474.
2. Kishwan J, Pandey R, Dadhwal VK. India's forest and tree cover: Contribution as a carbon sink. Technical Paper (Vol. 248006). Dehradun – 248006 (Uttarakhand) India. 2009;17. Available:[http://www.envfor.nic.in/mef/Technical\\_Paper/Pdf](http://www.envfor.nic.in/mef/Technical_Paper/Pdf) (Accessed 12/08/2017)
3. FAO. FRA-Global forest resources assessment 2015. Desk reference. Rome, Italy. 2015;253. Available:<https://doi.org/10.1002/2014GB005021/Pdf> (Accessed 9/04/2017)
4. Pandya IY, Salvi H, Chahar O, Vaghela N. Quantitative analysis on carbon storage of 25 valuable tree species of Gujarat, Incredible India. *Indian J. Sci. Res*. 2013;4(1):137–141.
5. Suryawanshi MN, Patel AR, Kale TS, Patil PR. Carbon sequestration potential of tree species in the environment of North Maharashtra University Campus, Jalgaon (MS) India. *Biosci. Discov*. 2014;5(2):175–179.
6. Potadar V, Patil S. Carbon storage and sequestration by trees in and around University Campus of Aurangabad City, Maharashtra. *Int. J. Innov. Res. Sci. Eng. Technol*. 2016;5(4):5459–5468.
7. Heineman KD, Turner BL, Dalling JW. Variation in wood nutrients along a tropical soil fertility gradient. *New Phytol*. 2016;16.
8. IPCC. Climate Change 2014 Synthesis Report. Contribution of Working Groups I, II and III to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change [Core Writing Team, R. K. Pachauri and L. A. Meyer (Eds.)], Geneva, Switzerland. 2014;169.
9. Mganga ND, Lyaruu HV, Banyikwa F. Above-ground carbon stock in a forest subjected to decadal frequent fires in

- western Tanzania. *J. Biodivers. Environ. Sci.* 2017;10(2):25–34.
10. Chave J, Andalo C, Brown S, Cairns MA, Chambers JQ, Eamus D, Yamakura T. Tree allometry and improved estimation of carbon stocks and balance in tropical forests. *Oecologia.* 2005;145(1):87–99.
  11. Zhang C, Ju W, Chen JM, Li D, Wang X, Fan W, Li M. Mapping forest stand age in China using remotely sensed forest height and observation data. *J. Geophys. Res. Biogeosciences.* 2014;119:1163–1179.
  12. Karki S, Joshi NR, Udas E, Adhikari MD, Sherpa S, Kotru R, Karky BS, Chettri N, Ning W. Assessment of forest carbon stock and carbon sequestration rates at the ICIMOD knowledge park in Godavari, Nepal. ICIMOD Working Paper 2016/6. Kathmandu: ICIMOD 2016. 2016;6:52.
  13. Vallet P, Meredieu C, Seynave I, Bélouard T, Dhôte JF. Species substitution for carbon storage: Sessile oak versus Corsican pine in France as a case study. *For. Ecol. Manage.* 2009;257(4):1314–1323.
  14. Vashum T, Jayakuma S. Methods to estimate above-ground biomass and carbon stock in natural forests - A review. *J. Ecosyst. Ecography.* 2012;2(4):7.
  15. Klein T, Hoch G. Tree carbon allocation dynamics determined using a carbon mass balance approach. *New Phytol.* 2015; 205(1):147–159.
  16. Clare C, Ford A, Free C, Hofmann C, Horvitz E, May E, Vara R. Carbon sequestration and its relationship to forest management and biomass harvesting in Vermont. Middlebury College. 2010;77. Available:<http://www.middlebury.edu/Pdf> (Accessed 12/12/2017)
  17. Wang YC. Tree species diversity and carbon storage in air quality enhancement zones in Taiwan. *Aerosol Air Qual. Res.* 2015;15(4):1291–1299.
  18. Hall JM, van Holt T, Daniels AE, Balthazar V, Lambin EF. Trade-offs between tree cover, carbon storage and floristic biodiversity in reforesting landscapes. *Landsc. Ecol.* 2012;27(8):1135–1147.
  19. Keenan RJ, Reams GA, Achard F, de Freitas JV, Grainger A, Lindquist E. Dynamics of global forest area: Results from the FAO global forest resources assessment 2015. *For. Ecol. Manage.* 2015;352:9–20.
  20. Patra B, Dey SK, Das MT. Forest management practices for conservation of biodiversity: An Indian perspective. *Int. J. Environ. Biol.* 2015;5(4):93–98.
  21. Strassmann KM, Joos F, Fischer G. Simulating effects of land use changes on carbon fluxes: Past contributions to atmospheric CO<sub>2</sub> increases and future commitments due to losses of terrestrial sink capacity. *Tellus.* 2008;60(4):583–603.
  22. Deng L, Bin Liu G, Ping Shangguan Z. Land-use conversion and changing soil carbon stocks in China's 'Grain-for-Green' program: A synthesis. *Glob. Chang. Biol.* 2014;20(11):3544–3556.
  23. Lal R. Soil carbon sequestration: Land and water use options for climate change adaptation and mitigation in agriculture. SOLAW Background Thematic Report – TRO4B. Rome, Italy. 2009;36. Available:<https://doi.org/10.1016/j.geoderma.2004.01.032/Pdf> (Accessed 10/06/2017)
  24. Xu X, Yang G, Tan Y, Tang X, Jiang H, Sun X, Li H. Impacts of land use changes on net ecosystem production in the Taihu Lake Basin of China from 1985 to 2010. *J. Geophys. Res. Biogeosciences.* 2017; 122(3):690–707.
  25. Baccini A, Walker W, Carvalho L, Farina M, Sulla-Menashe D, Houghton RA. Tropical forests are a net carbon source based on aboveground measurements of gain and loss. *Science.* 2017;358(6360): 230–234.
  26. Sundarapandian S, Swamy PS. Short-term population dynamics of tree species in tropical forests at Kodayar in the Western Ghats of Tamil Nadu, India. In Proceedings of the International Academy of Ecology and Environmental Sciences. 2013;3(3); 191–207.
  27. Thompson I, Mackey B, McNulty S, Mosseler A. Forest resilience, biodiversity, and climate change. A synthesis of the biodiversity/resilience/stability relationship in forest ecosystems. Montreal. 2009;43.
  28. Kimaro J, Lulandala L. Human influences on tree diversity and composition of a coastal forest ecosystem: The case of Ngumburuni Forest Reserve, Rufiji, Tanzania. *Int. J. For. Res.* 2013;1–7.
  29. Joyi O, Utanga MM, Dindi OO, Ynekulu JA, Ahman EAB. The effect of forest fragmentation on tree species abundance and diversity in the Eastern Arc Mountains of Tanzania. *Appl. Ecol. Environ. Res.* 2015;13(2):307–324.

30. del Río M, Barbeito I, Bravo-Oviedo A, Calama R, Cañellas I, Herrero C, Montero G, Moreno-Fernández D, Ruíz-Peinado R, Bravo F. Mediterranean pine forests: Management effects on carbon stocks. *Managing Forest Ecosystems: The Challenge of Climate Change*. 2017;31. Available:<https://doi.org/10.1007/978-3-319-28250-3/Pdf> (Accessed on 4/08/2017)
31. Williams M, Ryan CM, Rees RM, Sambane E, Fernando J, Grace J. Carbon sequestration and biodiversity of re-growing miombo woodlands in Mozambique. *For. Ecol. Manage.* 2008;254(2):145–155.
32. Mandal RA, Jha PK, Dutta IC, Thapa U, Karmacharya SB. Carbon sequestration in tropical and subtropical plant species in collaborative and community forests of Nepal. *Adv. Ecol.* 2016;2016:7.
33. Ostadhashemi R, Shahraji TR, Roehle H, Limaie SM. Estimation of biomass and carbon storage of tree plantations in northern Iran. *J. For. Sci.* 2014;60(9):363–371.
34. Ruiz-Jaen MC, Potvin C. Can we predict carbon stocks in tropical Ruiz-Jaen, M. C., & Potvin, C. Can we predict carbon stocks in tropical ecosystems from tree diversity? Comparing species and functional diversity in a plantation and a natural forest. *The new phytologist. New Phytol.* 2011;189(4): 978–87.
35. Jeyanny V, Husni MH, Wan Rasidah K, Siva Kumar B, Arifin A, Kamarul Hisham M, Hisham MK. Carbon stocks in different carbon pools of a tropical lowland forest and a montane forest with varying topography. *J. Trop. For. Sci.* 2014;26(4): 560–571.
36. Crossland M. The carbon sequestration potential of hedges managed for woodfuel. *The Organic Research Centre*. 2015;41. Available:<https://doi.org/10.1017/pdf> (Accessed 9/09/2017)
37. Powers JS, Becknell JM, Irving J, Prez-Aviles D. Diversity and structure of regenerating tropical dry forests in Costa Rica: Geographic patterns and environmental drivers. *For. Ecol. Manage.* 2009;258(6):959–970.
38. Wu H, Guo Z, Gao Q, Peng C. Distribution of soil inorganic carbon storage and its changes due to agricultural land use activity in China. *Agric. Ecosyst. Environ.* 2009;129(4):413–421.
39. Esquivel MJ, Harvey CA, Finegan B, Casanoves F, Skarpe C. Effects of pasture management on the natural regeneration of neotropical trees. *J. Appl. Ecol.* 2008;45(1):371–380.
40. Gao Y, Luo P, Wu N, Yi S, Chen H. Grazing intensity impacts on carbon sequestration in an alpine meadow on the Eastern Tibetan Plateau. *Res. J. Agric. Biol. Sci.* 2007;3(6):469–479.
41. MacFarlane DW, Kinzer AT, Banks JE. Coupled human-natural regeneration of indigenous coastal dry forest in Kenya. *For. Ecol. Manage.* 2015;354:149–159.
42. Briske DD, Derner JD, Milchunas DG, Tate KW. An evidence-based assessment of prescribed grazing practices. In *Conservation Benefits of Rangeland Practices: Assessment, Recommendations, and Knowledge Gaps*, USDA. 2011; 21–74.
43. Mugasha WA, Mwakalukwa EE, Luoga E, Malimbwi RE, Zahabu E, Silayo D, Sola G, Crete P, Henry M, Kashindye A. Allometric models for estimating tree volume and aboveground biomass in lowland forests of Tanzania. *Int. J. For. Res.* 2016;1–13.
44. Cheng J, Wu GL, Zhao LP, Li Y, Li W, Cheng JM. Cumulative effects of 20-year exclusion of livestock grazing on above- and belowground biomass of typical steppe communities in arid areas of the Loess Plateau, China. *Plant, Soil Environ.* 2011;57(1):40–44.
45. Kalaba FK, Quinn CH, Dougill AJ, Vinya R. Floristic composition, species diversity and carbon storage in charcoal and agriculture fallows and management implications in Miombo woodlands of Zambia. *For. Ecol. Manage.* 2013;304(2013):99–109.
46. Eneji IS, Obinna O, Azua ET. Sequestration and carbon storage potential of tropical forest reserve and tree species located within Benue State of Nigeria. *Journal of Geoscience and Environment Protection.* 2014;157–166.
47. Gilroy JJ, Woodcock P, Edwards FA, Wheeler C, Medina UCA, Hugaasen T, Edwards DP. Optimizing carbon storage and biodiversity protection in tropical agricultural landscapes. *Glob. Chang. Biol.* 2014;20(7):2162–2172.
48. Bradford JB, Jensen NR, Domke GM, D'Amato AW. Potential increases in natural disturbance rates could offset forest management impacts on ecosystem

- carbon stocks. *For. Ecol. Manage.* 2013;308:178–187.
49. Silayo DA, Tarimo MCT, Kweka ARE, Muganda AM. Impacts of human induced activities on species composition and diversity in Miombo woodlands of Bagamoyo district, Tanzania. *J. Korean Assoc. African Stud.* 2006;2:223–243.
  50. Axelsson C, Jones S, Haywood A, Suarez L, Wilkes P, Woodgate W, Mellora A. Key attributes for monitoring and assessment of Australian forests: A land management perspective. *CEUR Workshop Proc.* 2012;1328:9.
  51. Henry M, Bombelli A, Trotta C, Alessandrini A, Birigazzi L, Sola G, Saint-André L. GlobAllomeTree: International platform for tree allometric equations to support volume, biomass and carbon assessment. *IForest.* 2013;6:326-330. Available:<https://doi.org/10.3832/ifor0901-006.pdf> (Accessed on 10/04/ 2018)
  52. Shankar U. A case of high tree diversity in a sal (*Shorea robusta*) dominated lowland forest of Eastern Himalaya: Floristic composition, regeneration and conservation. *Curr. Sci.* 2001;81(7):776–786.
  53. Bharathi S, Prasad AGD. Regeneration status of tree composition in the sacred groves of Mercara, Central Western Ghats. *Eur. J. Exp. Biol.* 2015;5(10):29–40.
  54. Hitimana J, Kiyiapi JL, Njunge JT. Forest structure characteristics in disturbed and undisturbed sites of Mt. Elgon Moist Lower Montane Forest, Western Kenya. *For. Ecol. Manage.* 2004;194(1–3):269–291.
  55. Kauffman JB, Arifanti VB, Basuki I, Kurnianto S, Novita N, Murdiyarto D, Donato DC, Warren MW. Protocols for the measurement, monitoring, and reporting of structure, biomass, carbon stocks and greenhouse gas emissions in tropical peat swamp forests. Working Paper 221. Bogor, Indonesia: CIFOR. 2016;54.
  56. Ligate E, Chen C, Wu C. Evaluation of soil fertility status based on CEC and variation across disturbed and intact tropical coastal forests sites in Tanzania. *Asian J. Environ. Ecol.* 2018;6(2):1–12.
  57. URT. National forest resources monitoring and assessment of tanzania mainland; ministry of natural resources & tourism, tanzania forest services agency in Collaboration with the government of finland and food and agriculture Organization (FAO) of the United Nat. Dar es Salaam, Tanzania. 2015;124. Available:<http://www.fao.org/forestry/Pdf> (Accessed 12/04/2017)
  58. RGB, Chavan BL. Sequestered standing carbon stock in selective tree species grown in University campus at Aurangabad, Maharashtra, India. *Int. J. Eng. Sci. Technol.* 2010;2(7):3003–3007.
  59. Chen G, Shen H, Cao J, Zhang W. The influence of tree species on carbon storage in northern China. *For. Chron.* 2016;92(3): 316–321.
  60. Malimbwi R, Zahabu E. The analysis of sustainable charcoal production systems in Tanzania. Sokoine University of Agriculture, Morogoro, Tanzania. 2009;33. Available:<http://www.fao.org/docrep/012/i1321e/i1321e10.Pdf> (Accessed 9/06/2017)
  61. Mugasha W, Bollandas O. Allometric models for prediction of aboveground biomass of single trees in Miombo woodlands in Tanzania. *Clim. Chang. Impacts, Mitig. Adapt. Program. Sci. Conf.* 2012;8–17.
  62. IPCC. Good practice guidance for land use, land use change and forestry. (K. Jim Penman, Michael Gytarsky, Taka Hiraiishi, Thelma Krug, Dina Kruger, Riitta Pipatti, Leandro Buendia, Kyoko Miwa, Todd Ngara, (Eds.), Institute for Global Environmental Strategies. Hayama, Kanagawa Japan: IPCC National Greenhouse Gas Inventories Programme. Kanagawa, Japan. 2003;590.
  63. Feldpausch TR, et al. Height-diameter allometry of tropical forest trees. *Biogeosciences.* 2011;8(5):1081–1106.
  64. Schnitzer SA, Bongers F. Increasing liana abundance and biomass in tropical forests: Emerging patterns and putative mechanisms. *Ecol. Lett.* 2011;14(4):397–406.
  65. Hinkle ST. Biomass and carbon pools of a Miombo woodland ecosystem in Southern Tanzania. *Glob. J. Wood Sci. For. Wildl.* 2017;5(4):196–204.
  66. Pfeifer M, et al. Land use change and carbon fluxes in East Africa quantified using earth observation data and field measurements. *Environ. Conserv.* 2013; 40(3):241–252.
  67. Mligo C. The impact of livestock grazing on soil characteristics in Mount. *J. Geosci. Environ. Prot.* 2015;3:24–37.

68. Chazdon RL, Uriarte M. Natural regeneration in the context of large-scale forest and landscape restoration in the tropics. *Biotropica*. 2016;48(6):709–715.
69. Elmqvist T, Pyykönen M, Tengö M. Spontaneous regeneration of tropical dry forest in Madagascar: The social–ecological dimension. In *Reforesting Landscapes*. 2010;10:297–377.
70. Chazdon RL. Tropical forest regeneration. *Bol. do Mus. Para. Emílio Goeldi. Ciências Nat.* 2012;7(3):195–218.
71. Martin PA, Newton AC, Bullock JM. Carbon pools recover more quickly than plant biodiversity in tropical secondary forests. In *Proceedings of the Royal Society B: Biological Sciences*. 2013;280: 11.
72. Wheeler CE, Omeja PA, Chapman CA, Glipin M, Tumwesigye C, Lewis SL. Carbon sequestration and biodiversity following 18 years of active tropical forest restoration. *For. Ecol. Manage.* 2016;373: 44–55.
73. Pereira Júnior LR, de Andrade EM, de Q HA, Palácio PCL, Raymer JC, Ribeiro Filho, Pereira FJS. Carbon stocks in a tropical dry forest in Brazil. *Rev. Ciência Agrônômica*. 2016;47(1):32–40.
74. Condit R. Methods for estimating above-ground biomass of forest and replacement vegetation in the tropics. Center for Tropical Forest Science Research Manual. 2008;73.
75. Hlásny T, Barcza Z, Fabrika M, Balázs B, Churkina G, Pajtík J, Turčáni M. Climate change impacts on growth and carbon balance of forests in Central Europe. *Climate Research*. 2011;47(3):219–236.
76. Perrings C. Biodiversity, ecosystem services, and climate change the economic problem (Environmental Economics Series No. 120). Washington DC, USA. 2010; 45.
77. Oakes LE, Hennon PE, O’Hara KL, Dirzo R. Long-term vegetation changes in a temperate forest impacted by climate change. *Ecosphere*. 2014;5(10):135.
78. Hunt CAG. Carbon sinks and climate change: Forests in the fight against global warming. GB: Edward Elgar Publishing Ltd. 2009;261-262.
79. FAO. Biomass. Assessment of the status of the development of the standards for the terrestrial essential climate variables. Global Terrestrial Observing System, Rome, Italy, Version 10, 30. 2009;86.